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# **Environment & Sustainability Standing Committee March 3, 2011**

TO:

Chair and Members of Environment and Sustainability Standing

Committee

SUBMITTED BY:

Phillip Townsend, Director, Infrastructure and Asset Management

DATE:

February 15, 2011

SUBJECT:

Lake Water Quality

### **INFORMATION REPORT**

Allanami !

### **ORIGIN**

This report originates from staff.

#### BACKGROUND

During 2010, Halifax Regional Municipality (HRM) staff, and consultants under contract with HRM, collected water samples from 69 sample sites (62 in lakes and 7 in rivers) during the spring, summer and fall seasons (Attachment 3). These samples were tested for a variety of substances including Total Phosphorus (TP), which is typically a limiting nutrient for aquatic plant growth in freshwater systems. Although the TP results for the fall season were relatively low, those for spring and summer 2010 were high, resulting in high annual average concentration. These high TP results prompted a review of previous seasonal and annual TP results, revealing that they too have been high.

These high TP results are of concern for a few reasons but primarily because they signal a trend towards eutrophication, which typically means a reduction in amenity value to humans (Attachment 1). This trend is contrary to the intent of Policy E-18 in the RMPS, which identifies one of the two main goals of the Water Quality Monitoring Functional Plan as "... to stem the decline of lakes from the accelerated process of eutrophication, and sedimentation and inputs from other urban runoff by managing development on a watershed basis."

"Eutrophication" is the process by which lakes and rivers gain plant matter, and it is described according to a series of categories called "trophic states". Trophic states are measures along an absolute trophic scale, which describes the biological condition of a water body (Table 1 below).

Table 1 - Total phosphorus trigger ranges for Canadian lakes and rivers:

Trophic State	Canadian Trigger Ranges Trophic Status Total phosphorus (µg/L)
Ultra-oligotrophic	< 4
Oligotrophic	4-10
Mesotrophic	10-20
Meso-eutrophic	20-35
Eutrophic	35-100
Hyper-eutrophic	> 100

Based on the CCME Canadian Guidance Framework for the Management of Freshwater Systems (attachment 2).

Although related, 'trophic states' are not synonymous with water quality. Water quality is a term used to describe the condition of a water body in relation to human needs or values. Quality is not an absolute; the terms "good" or "poor" water quality only have meaning relative to the use of the water and the attitude of the user. An oligotrophic lake might have good water quality for swimming but be considered poor water quality for bass fishing.

Using the annual average TP concentration, most lakes tested in HRM in 2010 fall into the Meso-eutrophic classification - more so than any other year for which we have data (Table 2).

Table 2 - Summary of Annual Average TP results since 2006 (n=69):

Year	# Sites Meso-eutrophic or above (> 20 μg/L)	% Sites Meso-eutrophic or above (> 20 μg/L)	
2010	47	68	
2009	17	25	
2008	20	29	
2007	6	9	
2006	7	10	

Although HRM has not established consistent water quality objectives or standards that apply to all lakes, Harbour East and North West Community Councils have established management thresholds that are in force through approved Master Plans, Secondary Planning Strategies and associated Development Agreements. Table 3 presents Total Phosphorus thresholds and 2010 annual average TP results for affected lakes.

Table 3 - Summary of water quality management thresholds and 2010 TP results by lake:

Lake	Community Council	Management threshold (TP μg/L)	2010 Avg. TP results (μg/L)
Morris	HECC	15	17 – North; 16 – South
Russell	HECC	15	5
Kearney	NWCC	10	7
Paper Mill	NWCC	10	7

#### **DISCUSSION**

As mentioned in the background, high TP levels in HRM lakes are of concern. To meaningfully interpret and act on these results, staff needs to confirm the validity and reliability of the data and, upon confirmation, determine the source of the nutrient and identify management options for consideration.

Data validity was assessed via several procedures:

- spot-checking the accuracy of data used for analysis (comparison with results submitted by labs);
- comparing timing of high TP lake counts to changes in analytical laboratories & procedures used by those labs; and
- comparing data reporting methodology for differences due to reportable detection limit calculations.

The data used to reach these results appear to be fully valid, at least in accordance with the procedures used to date. Staff will conduct further analyses on the advice of external experts to vet these results.

Given confidence in the data, it does indicate a real change in the condition of several studied lakes. It is appropriate to expect that TP levels are also increasing in other lakes in HRM that are not subject to ongoing monitoring.

The reasons for the increase in TP levels are not yet known, but there are two possible mechanisms:

- 1) phosphorous entering from outside the lake body (i.e. from lakes, rivers and pipes upstream or overland flow from the surrounding watershed); and
- 2) phosphorous being reintroduced to the water column from bottom sediments due to natural biological, chemical and/or physical processes.

Potential sources of phosphorous from outside the lake are more obvious than those from inside, and they include land-clearing activity through development, landscaping practices and contributions from stormwater and/or wastewater systems. Staff are working on the Stormwater Management Functional Plan (SU-28 of the Regional Plan), which will address these issues and several others.

The trophic status of lakes reported here is based exclusively on Total Phosphorous – as endorsed by Environment Canada's Canadian Guidance Framework for the Management of Phosphorus in Freshwater systems. Other systems for trophic status evaluation do exist – for example the Trophic Status Index (TSI) and the Lake Evaluation Index (LEI); their suitability to the lakes in HRM will be assessed and future water quality reports will reflect the results of this assessment. Consultation will be sought with local freshwater experts (including those in provincial and federal agencies and local universities) to identify further analytical techniques and insights on the possible cause of these findings and associated remedies.

During 2010, several lakes were also observed to have significant aquatic plant growth, in excess of growth in recent years. As reported in a briefing note to this committee in February 2011, HRM is working with a community group to assess the potential to mitigate weed impacts on Lakes Banook and Micmac for 2011. Results from the lake water quality studies will be used to inform aquatic plant management considerations.

#### **BUDGET IMPLICATIONS**

There are no budget implications associated with this report. Options identified to address this issue through further study may have budget implications.

#### FINANCIAL MANAGEMENT POLICIES / BUSINESS PLAN

This report complies with the Municipality's Multi-Year Financial Strategy, the approved Operating, Project and Reserve budgets, policies and procedures regarding withdrawals from the utilization of Project and Operating reserves, as well as any relevant legislation.

#### **COMMUNITY ENGAGEMENT**

No community engagement occurred in relation to this report. As staff become more certain about the cause of phosphorus loading in the lakes, appropriate community engagement strategies will be identified and pursued.

#### **ATTACHMENTS**

- 1. Summary of Problems with Eutrophication (Table 2.2)
- 2. CCME Phosphorus Guidelines
- 3. Water Quality Sample Sites 2010

A copy of this report can be obtained online at http://www.halifax.ca/commcoun/cc.html then choose the appropriate Community Council and meeting date, or by contacting the Office of the Municipal Clerk at 490-4210, or Fax 490-4208.

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Table 2.2 Summary of the problems associated with eutrophication of lakes, reservoirs, and impoundments (modified from GEMS 1990).

	Caused or indirectly dependent on				
Quality of life concerns	Algal blooms and species composition	Excessive macrophyte and littoral algal growth	Altered thermal conditions	Mineral turbidity	Low dissolved solids
Water quality impairment					
<ul> <li>Taste and odour, colour, filtration, flocculation, sedimentation and other treatment difficulties</li> </ul>	Very frequent	Occasional	Occasional	Occasional	
<ul> <li>Hypolimnetic oxygen depletion; Fe, Mn, CO<sub>2</sub>, NH<sup>+</sup>4, CH<sub>4</sub>, H<sub>2</sub>S, etc.</li> </ul>	Frequent	Occasional	Occasional		
<ul> <li>Corrosion problems in pipes and other man-made structures</li> </ul>	Frequent	Occasional		Occasional	Occasional
Recreational impairment					
<ul><li>Poor aesthetics</li><li>Hazard to bathers</li></ul>	Frequent	Occasional Frequent		Occasional	
<ul> <li>Increased health hazards</li> <li>Fisheries impairment</li> </ul>	Occasional	Occasional	Occasional		
<ul> <li>Fish mortality</li> </ul>	Occasional		Occasional		
<ul> <li>Undesirable fish stocks</li> </ul>	Frequent	Frequent	Occasional	Occasional	
Ecological impairment Aging and reduced holding capacity and flow	Very frequent	Frequent	Frequent	Occasional	Occasional
<ul><li>Silting</li><li>Pipe and screen</li></ul>	Occasional	Occasional Occasional		Frequent	
clogging					



### Canadian Water Quality Guidelines for the Protection of Aquatic Life

### PHOSPHORUS: CANADIAN GUIDANCE FRAMEWORK FOR THE MANAGEMENT OF FRESHWATER SYSTEMS

hosphorus (P; CAS No. 7723-14-0, atomic mass 30.97376) is a highly reactive, multivalent, nonmetal of the nitrogen group in the period table that is never found free in nature. Phosphate rock, which contains the mineral apatite, an impure tri-calcium phosphate, is an important source of phosphorus. Phosphorus is an essential nutrient for all living organisms; living matter contains about 0.3 percent dry weight phosphorus (Horne and Goldman 1994). It plays a major role in biological metabolism, and when compared to other macronutrients required by biota, phosphorus is the least abundant and commonly the first nutrient to limit biological productivity (Wetzel 2001). Water bodies containing low phosphorus concentrations (i.e., unimpacted sites) typically support relatively diverse and abundant aquatic life that are self-sustaining and support various water uses. However, elevated phosphorus concentrations can adversely affect aquatic ecosystems (Chambers et al. 2001).

The phosphorous guideline follows a framework-based approach that accommodates the non-toxic endpoints associated with phosphorus and can be incorporated into existing management strategies. The framework offers a tiered approach where phosphorus concentrations should not (i) exceed predefined 'trigger ranges'; and (ii) increase more than 50% over the baseline (reference) levels. The trigger ranges are based on the range of phosphorus concentrations in water that define the reference trophic status for a site. If the upper limit of the range is exceeded, or is likely to be exceeded, further assessment is required. When assessment suggests the likelihood of undesired change in the system, a management decision must be made.

#### Phosphorus in the Aquatic Environment

Phosphorus in aquatic systems occurs in three forms: inorganic phosphorus, particulate organic phosphorus, and dissolved (soluble) organic phosphorus. Aquatic plants require inorganic phosphate for nutrition, typically in the form of orthophosphate ions (PO<sub>4</sub><sup>3</sup>). This is the most significant form of inorganic phosphorus, and is the only form of soluble inorganic phosphorus directly utilized by aquatic biota. This form of phosphate is transferred to consumers and decomposers as organic

phosphate. Most of the phosphorus (up to 95%) in fresh water occurs as organic phosphates, cellular constituents of organisms, and within or adsorbed to inorganic and dead particulate organic matter (Wetzel 2001). This is subsequently made available for recycling via mineralization and decomposition.

Phosphate concentrations tend to increase with increases in total phosphorus (TP), but the proportion of phosphate declines with increasing TP (Hudson et al. 2000). The turnover rate of PO<sub>4</sub><sup>3-</sup> in phosphorus limited systems is extremely rapid, thus making its measurement difficult. Filtration of the water sample prior to analysis for PO<sub>4</sub><sup>3-</sup> can overestimate biologically available phosphorus (Fisher and Lean 1992) and conventional methods for measuring PO<sub>4</sub><sup>3-</sup> generally overestimate phosphate concentrations (Hudson et al. 2000). Based on these limitations, TP is generally recommended as a meaningful measurement of phosphorus in surface waters (Wetzel 2001).

In the majority of lakes, phosphorus is normally the limiting nutrient for algal growth. However, in some areas (prairie lakes and rivers) nitrogen is increasingly being found to be the most important nutrient. TP concentrations in non-polluted natural waters extend over a very wide range from <1  $\mu g \cdot L^{-1}$  in ultra-oligotrophic waters, to >200  $\mu g \cdot L^{-1}$  in highly eutrophic waters; however, most uncontaminated freshwaters contain between 10 and 50  $\mu g \cdot L^{-1}$  of TP (Wetzel 2001). Phosphorus levels of freshwaters are generally lowest in mountainous regions of bedrock geomorphology (e.g., Boreal and Pre-Cambrian Shields), and increase in lowland waters derived from sedimentary rock deposits (e.g., the Boreal Plains of Alberta). Lakes rich in organic matter, and bogs, tend to exhibit high TP concentrations.

Sedimentation of particulate phosphorus results in a slow but continuous loss from the water column. Phosphate is precipitated as insoluble iron, calcium, or aluminium phosphate and then released slowly. Exchange of phosphorus across the sediment/water interface is regulated by oxidation-reduction (redox) interactions, which are dependent on oxygen supply, mineral solubility and sorptive mechanisms (Stumm and Morgan 1996), the metabolic activities of bacteria and fungi, and turbulence from physical and biotic activities (Wetzel 2001). Lake

sediments contain much higher concentrations of phosphorus than water. Under aerobic conditions, the exchange equilibria are largely unidirectional toward the sediments; however, under anaerobic conditions, inorganic exchange at the sediment-water interface is strongly influenced by redox conditions. Sediment phosphorus release (internal loading) can be an important source of phosphorus and can maintain high phosphorus concentrations in the water column, even in the absence of significant external loading (Marsden 1989; Holz and Hoagland 1999).

The first response of an aquatic system to phosphorus additions is increased plant and algal productivity and biomass. Although this may be desirable in some cases, beyond a certain point, further phosphorus additions to phosphorus-limited systems can cause undesirable effects, such as: (i) decrease in biodiversity and changes in dominant biota; (ii) decline in ecologically sensitive species and increase in tolerant species; (iii) increase in plant and animal biomass; (iv) increase in turbidity; (v) increase in organic matter, leading to high sedimentation; and (vi) anoxic conditions (Mason 1991; Environment Canada 2004). When the excessive plant growth includes certain species of cyanobacteria, toxins may be produced, causing increased risk to aquatic life, livestock, and human health (Chambers et al. 2001).

The potential human quality of life concerns that may relate to eutrophication are: (i) treatment of potable water may be difficult and costly; (ii) the supply may have an unacceptable taste or odour problem; (iii) the water may be injurious to health; (iv) the aesthetic/ recreational value of the water may decrease; (v) macrophyte growth may impede water flow and navigation; (vi) commercially important species (such as salmonoids and coregonids) may disappear (Mason 1991; Environment Canada 2004).

#### Water Quality Guideline Derivation

Currently, no national environmental quality guidelines exist for phosphorus, although individual provinces may have guidelines or objectives (Environment Canada 2004). The Protocol for the Derivation of Guidelines for the Protection of Aquatic Life (CCME 1991) is intended to deal specifically with toxic substances, and provide numerical limits or narrative statements based on the most current, scientifically defensible toxicological data. Phosphorus does not fit this model because it is non-toxic to aquatic organisms at levels and forms present in the environment; however, secondary effects, such as eutrophication and oxygen depletion are serious concerns. Because aquatic communities are generally adapted to

ambient conditions, it is neither feasible nor desirable to establish a single guideline value for phosphorus. Some of the effects of phosphorus are aesthetic and thus include an element of subjectivity. What is considered nuisance plant growth to some may be desirable to others. Based on these realities, guidelines for phosphorus are not derived; rather a guidance framework (Figure 1) that is consistent with CCME guideline principles is developed. The framework accommodates the non-toxic endpoints associated with phosphorus loading and permits site-specific management of phosphorus.

The framework provides a tiered approach where water bodies are marked for further assessment by comparing their trophic status to predefined 'trigger ranges'. The trigger ranges are based on a range of phosphorus concentrations in water that define the reference status for a site. Using such a scheme, sites with similar characteristics are classified together irrespective of whether they might possess these features naturally or as a result of human influence. Since the reference condition for the water body in question is defined at the onset of the framework, this problem is readily overcome; predefined states, whether determined through hindcasting or by using best available data, are always used to set the trigger range.

#### Steps in Guidance Framework

#### Set Ecosystem Goals and Objectives

As a first step, it is crucial to set ecosystem goals and objectives (e.g., enhance, protect, or restore). The objective can be set for a healthy aquatic ecosystem, with the goals being unimpaired human uses, and a diverse and functioning aquatic ecosystem. By setting the diverse aquatic ecosystem as a goal, a desire to prevent the loss of species is established. Similarly, the desire for a functioning aquatic ecosystem recognizes that ecosystems do things that are both inherently of value and of value for human uses and desires. These objectives are important because they will guide management decisions made later.

#### Define Reference/ Baseline Conditions

Establishing the reference condition is the most important step in the framework because it determines the trigger range that is used for comparison. In some cases, historical data may be available, but in most cases there will be a need to estimate reference (baseline) phosphorus

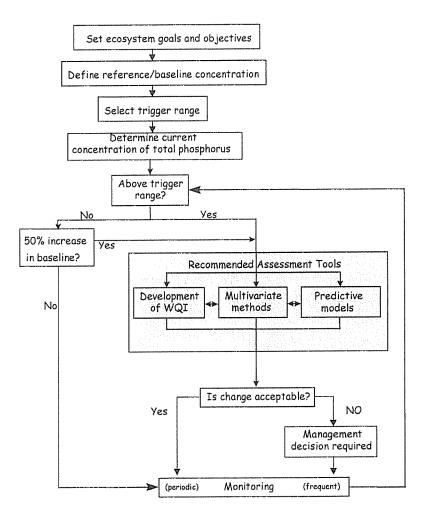


Figure 1. Canadian Guidance Framework for the management of phosphorus in freshwater systems.

concentrations. Several options are available for this, ranging from use of available historical data to derivation and application of predictive models to hindcast predevelopment phosphorus values (Environment Canada 2004). Many jurisdictions (e.g., British Columbia and Ontario) which are actively managing phosphorus, have already established reference conditions that could be used in the framework. In addition, reference conditions will be relatively simple to determine in areas with little or no development. In geographical areas where there is a high density of water bodies, a single reference condition may be established for the entire area.

#### Select Trigger Ranges

Australia, New Zealand (NWQMS 1999), and the USEPA (EPA 2000) consider ecosystem classification in setting their nutrient guidelines. In the Canadian framework, trigger ranges are based on the trophic classification of the baseline condition or the status of reference sites.

Internationally accepted OECD (Organisation for Economic Co-operation and Development) trophic status values (Vollenweider and Kerekes 1982) are the

recommended trigger ranges (Table 1). The only proposed variation is that the OECD meso-eutrophic category (10-35  $\mu g \cdot L^{-1}$ ) is subdivided into mesotrophic (10-20  $\mu g \cdot L^{-1}$ ) and meso-eutrophic (20-35  $\mu g \cdot L^{-1}$ ). This subdivision was necessary because considerable variation in community composition and biomass exist in Canadian waters over the OECD range of 10-35  $\mu g \cdot L^{-1}$ . These trigger ranges are recommended for both rivers and lakes.

A trigger range is a desired concentration range for phosphorus; if the upper limit of the range is exceeded, it indicates a potential environmental problem, and therefore, "triggers" further investigations. Natural physical and chemical water quality variables (e.g., salinity, pH, nutrients) inherently vary within and between ecosystem types, and so the preferred method for determining the trigger ranges is to use similar, high quality reference sites to determine natural levels. These ranges are then categorised according to the trophic status of the reference site (Table 1). This approach provides a trigger range that is relevant to the ecosystem type and locality. These phosphorus limits allow management to define where their water bodies lie, and define a trigger range for that water body.

Table 1. Total phosphorus trigger ranges for Canadian lakes and rivers.

Trophic Status	Canadian Trigger Ranges Total phosphorus (µg·L <sup>-1</sup> )
Ultra-oligotrophic	< 4
Oligotrophic	4-10
Mesotrophic	10-20
Meso-eutrophic	20-35
Eutrophic	35-100
Hyper-eutrophic	> 100

The selection of appropriate trigger ranges and reference conditions can potentially benefit from the development and application of an ecoregional approach (Environment Canada 2004). Ecoregions provide a means of classifying ecologically distinct areas, where each region can be viewed as a discrete system made up of areas of similar geographical landform, soil, vegetation, climate, wildlife, water, etc. The use of ecoregions can improve predictability of nutrient enrichment effects. They can help differentiate between natural and anthropogenic contributions to nutrient enrichment, reduce variability in trigger ranges within a class and among classes, and contribute to improved assessment and development of trigger ranges.

#### Determine Current Phosphorus Concentration

Under normal conditions, TP is the only meaningful measurement of phosphorus for water (Wetzel 2001). TP can be expressed as a single measurement taken at spring turnover or as an average of several observations made on a seasonal basis; it may be an estimate for a specific zone (e.g., euphotic zone), or as a whole lake approximation. It is important that an appropriate number of samples are collected to accurately reflect TP concentrations in a system. Specific attention should be given to sites that are receiving variable phosphorus loads or exhibiting marked morphological and hydrological differences (Environment Canada 2004).

## Compare Current or Predicted Concentration to Trigger Range

The upper concentration of the trigger range represents the maximum acceptable concentration of phosphorus within each of the trophic categories. If the upper limit of the trigger range is exceeded, or is likely to be exceeded, there is a risk of an impact either occurring or having occurred. At this stage, additional information on local environmental factors needs to be considered, and thus further assessment is recommended. The assessment could potentially lead to remedial advice and the restoration of a degraded water body. If the trigger range is not exceeded, the risk of an impact is regarded as low.

## Compare Current or Predicted Concentration to Baseline Condition

Due to the general nature of the trigger values and the size of some of the phosphorus ranges defined, a second precaution is taken in the assessment of possible effects of phosphorus. In the event that the trigger value has not been exceeded, the question is now raised as to the degree of increase in phosphorus levels from the baseline. Up to a 50% increase in phosphorus concentrations above the baseline level is deemed acceptable (OMOE 1997). In large lakes, the 50% increase should be applied to the most sensitive areas (e.g., river mouth, point sources, or the littoral zone) rather than averaged over the whole lake. The 50% increase check is also applied to river systems. It is important to recognize that the 50% increase limit in lakes that already have high phosphorus baseline (up to 12  $\mu g \cdot L^{-1}$ ) may not protect against decreases in dissolved oxygen. However, in the absence of empirical data to recommend an alternative, the 50% increase limit is deemed preferable to no limit. If a 50% increase from

## Canadian Water Quality Guidelines for the Protection of Aquatic Life

the baseline is not observed, then there is considered a low risk of adverse effects, and only monitoring is required for these sites. If the increase from the baseline is greater than 50%, the risk of observable effects is considered to be high, and further assessment is recommended.

#### Recommended Assessment Tools

The recommended assessment options currently in use in Canada, the US, and other parts of the world are presented in detail in Environment Canada (2004). In summary, the tools fall into the following three categories:

- i. A water quality index can be used as a surrogate for phosphorus as it can provide a single value that identifies the current state of the ecosystem, including the percent change from the reference/baseline condition (e.g., Johnes et al. 1994).
- Multivariate methods can also assist in comparing current conditions to baseline conditions, and the degree of impairment can be identified both spatially and temporally (Reynoldson and Day 1998; Kilgour et al. 1998; OMOE 1999).
- iii. Predictive models can be applied for the management of phosphorus. For example, Lakeshore Capacity (Dillon and Rigler 1975) and paleoecology based phosphorus reconstruction (Dixit et al. 1999) models have been successfully used to estimate baseline phosphorus concentrations and to assess the magnitude and the rate of change.

These assessment tools should be viewed with the caveat that many of them were developed for specific water types or for specific areas with underlying topographic and geological assumptions. Although many of these methods can be adapted to the specific user's situation/need, care must be taken in selecting an approach that is both technically feasible and realistic to the user's needs. Furthermore, the application of any assessment tool in defining phosphorus concentration may not be exclusive, and it may be necessary to adapt a combination of approaches.

#### Management Decisions

Once the potential increases in phosphorus concentrations have been assessed, the results are compared to the original goals (i.e., reference or baseline conditions) set at the beginning of the framework. The degree of change is then assessed on a management level, and the question raised. "are these changes acceptable?".

Management decisions are a critical step in the framework that links back to the objectives and outcomes of the program via monitoring. Management of phosphorus should include both short-term management strategies and options, which primarily focus on operational activities; and long-term management strategies, which focus on nutrient reduction, flow management, education, monitoring, and research.

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# HRM Water Quality Sampling Program List of Sampled Lakes & Rivers 2009

Albro Lake

Barrett Lake

Bell Lake

Bissett Lake

Black Duck Pond

Black Point Lake

Chocolate Lake

Cranberry Lake

De Said Lake

Dents Punch Bowl

Drain Lake

Duck Lake

Fenerty Lake

First Chain Lake

First Lake

Fletchers Lake

Frenchman Lake

Frog Pond

Governor Lake

Grand Lake (Shubenacadie)

Half Mile Lake

Hubley Big Lake

Kearney Lake

Kidston Lake

Kinsac Lake

Lake Banook

Lake Charles

Lake Echo

Lake Micmac

Lake Thomas - North

Lake Thomas - South

Lake William

Lisle Lake

Little Albro Lake

Long Lake

Long Pond

Loon Lake

Lovett Lake

MacIntosh Run at Mouth

MacIntosh Run at Roaches Pond (Bacteria only)

Maynard Lake

McCabe Lake

Moody Lake

Morris Lake (North Basin)

Morris Lake (South Basin)

Nine Mile River at Highway 103

Nine Mile River at Mouth

Oathill Lake

Paper Mill Lake

Penhorn Lake

Powder Mill Lake

Red Bridge Pond

Rocky Lake

Russell Lake

Sackville River

Sackville River at Highway 102 (Bacteria only)

Sandy Lake (Bedford)

Sandy Lake (Glen Arbour)

Second Lake

Settle Lake

Sheldrake Lake

Smelt Brook

Springfield Lake

Stillwater Lake

The Mill Pond (Three Mile Pond)

Third Lake

Tucker Lake

Upper Porters Lake

Whimsical Lake

Williams Lake

Winder (Whynder) Lake